

Bird Communities on Recently-logged and Mature Non-Industrial Private Forestlands in Virginia

Todd S. Fredericksen¹

School of Natural Sciences and Mathematics
Ferrum College
Ferrum, Virginia 24088

ABSTRACT

Most studies of bird community responses to logging in the eastern United States have occurred on government or industrial forestlands. Non-industrial private forestlands (NIPFs), however, represent the single largest forestland ownership category in the nation and the largest source of timber, particularly in eastern states. This study compares bird species abundance, richness, and composition on recently harvested NIPF stands with mature forest stands over three years during the early breeding season and peak spring migration in Virginia. A total of 79 bird species was observed on the study sites, 76 of which were observed on recently logged stands and 37 on control stands. Logged stands had approximately twice the mean bird abundance and 50% higher species richness, on average, than that of mature forest stands. Bird species abundance was negatively correlated with overstory (>10 m) tree cover and midstory (2-10 m) tree cover and positively correlated with herbaceous layer cover (<0.5 m) and the amount of large woody debris. Bird species abundance was not significantly correlated with shrub cover (0.5-2 m) or the abundance of snags. Relationships for species richness were similar except that species richness was not significantly correlated with overstory cover. Differences in species composition were evident among bird communities on logged and mature forest stands and were also affected by the intensity of logging.

Key words: biodiversity, forest management, logging, timber harvesting.

INTRODUCTION

Timber harvesting, often referred to as logging, can have significant impacts on wildlife species. Each species of wildlife has specific habitat requirements for food, water, cover, and nesting sites, and habitat conditions may be changed by timber harvesting. For example, logging will decrease the amount of overstory tree cover, but typically increases understory ground cover, understory food sources, and the amount of woody debris on the forest floor (Fredericksen et al., 2000; Brawn et al., 2001).

The response of birds to logging is of particular interest to many conservation biologists and natural historians. Bird species are useful indicators of habitat change (Balda, 1975). In addition, populations of many bird species have experienced steep declines in recent decades. There are particularly strong concerns about Neotropical migrants that breed in North American

forests (Robbins et al., 1989; Peterjohn et al., 1995) and whether logging may negatively affect forest interior bird species, particularly those in already fragmented landscapes (Robinson & Robinson, 1999). On the other hand, many bird species that require early successional forest habitat are also in decline (Brawn et al., 2001; Dettmers, 2003; Sauer et al., 2005).

Results of studies exploring the impacts of logging and other silvicultural treatments have been mixed and highly species-specific (Sallabanks & Arnett, 2005). Most of our knowledge of how birds respond to timber harvesting comes from research on industrial forestlands or government-owned forestland. Much less is known about how birds and other wildlife species respond to logging on non-industrial private forestlands (NIPFs), which comprise the majority of all forestland in the eastern United States (Birch, 1996) and from which the majority of timber is harvested in the United States (Best & Wayburn, 2001). These properties are often small in size and are imbedded within fragmented landscapes (Best & Wayburn, 2001; Butler &

¹email: tfredericksen@ferrum.edu

Leatherberry, 2004). NIPFs are diverse and include forest stands within residential properties, farms, and hunting club lands.

As in most eastern states, the majority of all forestland in Virginia is owned by non-industrial forest landowners (Thompson & Johnson, 1996). In Virginia, the mean size of these forest tracts is 90 ha, but it is highly skewed towards small tracts, with 62% being less than 40 ha (Thompson & Johnson, 1996). This report describes the results of a three-year study to determine the impacts of logging and logging intensity on the early breeding season/spring migration abundance, species richness, and species composition of bird species at the interface between the southern Blue Ridge and Upper Piedmont physiographic provinces of Virginia. Bird species abundance and richness were also correlated with habitat variables including vegetation cover at different levels within the forest, coarse woody debris, and snags.

METHODS

Study Sites

A total of 18 forest stands were identified in each year of the study with harvest intensities ranging from no harvesting (100% of original overstory cover) to clearcutting (0% overstory cover) (Table 1). In 2004, eight recently-harvested stands (harvested within the past two years) and 10 mature forest stands were used in the study. In 2005 and 2006, there were nine harvested stands and nine mature forest stands that had no recent history of logging. Most of the same stands were used in all three years of the study. Care was taken to make sure that the forest stands were similar with respect to pre-harvest stand composition. Stands were dominated by naturally-regenerated hardwood tree species with some scattered White Pine (*Pinus strobus*) and Virginia Pine (*Pinus virginiana*). They were similar in size (12-40 ha) and landscape matrix (percentage of farms, fields, urban development, and forest).

Field Sampling

Bird species richness and abundance were estimated using point counts with a 40-m radius. There were four survey points in each stand, with each point separated by 100 m. Points were located systematically along a transect approximately through the center of each stand. All birds heard or seen within the point count area were recorded and identified to species. Counts were conducted from 0600-1000 h, and 2-4 stands were sampled each day. Point counts were not made during periods of rain or windy conditions. All points on each

site were sampled twice between late-April and mid-May each year from 2004-2006, coinciding with the peak of bird migration through the area and the peak of bird song activity for most species as they establish territories on breeding grounds. This period of the year was considered to be the best time to sample bird density because they are more vocal when establishing territories and, hence, are more easily detected, and because the use of sites by spring migrants can also be evaluated. Data were summed over all census points for each census day and the average bird abundance derived from the average of the two census days in each year. Bird abundance was therefore expressed as mean abundance per sampling day per unit area. During the summer, another study examining mammal and herpetofaunal abundance allowed for additional presence/absence data to better estimate bird species richness during the growing season. Since the amount of time spent in each stand for all sampling activities (sampling of all wildlife groups) was approximately equivalent, bird species richness was derived from a list of all species heard or seen on the site throughout the entire sampling season, not just from those observed during point counts.

Habitat Sampling

Vegetation cover data were collected at 10 systematically random locations in each stand. Percentage overstory (>10 m tall) tree cover and midstory (2-10 m tall) cover were estimated using a transparent grid densiometer. Percentage shrub (0.5-2 m tall) cover and ground (<0.5 m tall) cover were estimated ocularly to the nearest 5% using a 1-m² sampling frame. At each sampling point, a prism count was conducted to estimate total live and dead (snag) tree basal area. Woody debris cover >5 cm in diameter was estimated using approximately ten 50-cm line-intercept transects within which the diameter of each woody debris item was measured. An index of woody debris was constructed using the sum of all woody debris diameters along transects.

Statistical Analyses

For species richness, species were counted if they were observed during formal sampling or if they were incidentally observed using the site during any visit to the site. The time spent on each site was roughly equivalent for all study sites allowing for comparisons of total species richness among sites. Analysis of variance was used to test the hypothesis of no difference between recently-logged and mature forest stands with respect to bird abundance or species

Table 1. Basal area and logging history of recently-logged and mature forest* stands in Franklin, Patrick, and Henry counties, Virginia.

Study Stand	Year of previous logging	Basal area m ² ha ⁻¹	Years sampled	Type of logging
Baker	2004	10.0	2005-6	Patch clearcutting/selective
Bowling	1960s	30.0	2004-6	Mature forest
Brubaker	2004	0.0	2005-6	Clearcut
Compton	2003	0.0	2004	Clearcut with site preparation
Ferrum Mtn.	1970s	24.5	2004-6	Mature forest
Ferrum Trail	1970s	26.5	2004-6	Mature forest
Ferrum Ridge	1970s	33.7	2004-6	Mature forest
Fredericksen	1960s	26.5	2004-6	Mature forest
Grice	2003	25.5	2004-6	Moderately selective
GT Lester	2003	2.7	2004-6	Heavy selective
Heck	1960s	36.2	2004-6	Mature forest
Hutchinson	1960s	30.0	2004	Mature forest
Kings Mountain	2003	5.0	2004-6	Heavy selective
Kitterman	1960s	40.0	2004-6	Mature forest
Naff	2003	23.8	2004-6	Moderately selective
Pettigo	1960s	32.2	2004-6	Mature forest
Potter	2002	17.0	2004	Moderately selective
Rocky Mt. 1	2003	16.5	2004-6	Heavy selective
Rocky Mt. 2	2004	10.0	2005-6	Heavy selective
Snow Creek	1960s	26.3	2004-6	Mature forest
Wagner	2003-4	12.0	2004-6	Heavy selective

*Indicates mature forest stand although light selective logging had occurred in some stands in the past (>30 years ago).

richness. Sampling year was included in the model. The natural logarithm of bird abundance was used because of unequal variances for bird abundance on recently-logged and mature forest stands. Pearson's correlation coefficient was used to test for the correlations between bird abundance and species richness with habitat variables. Differences were considered statistically different at $p \leq 0.05$, although p values ≤ 0.15 are reported to note trends towards statistical significance. Detrended correspondence analysis (DCA) was used to compare stands and stand-species associations with respect to bird species composition (PC-ORD, version 5, MJM software, Glenden Beach, Oregon, USA).

RESULTS

Habitat Assessment

Recently-logged stands had nearly three times the percentage herbaceous layer cover and amount of large woody debris than mature forest stands (Table 2), while mature forest stands had three times the percentage overstory cover and twice the percentage midstory cover as logged stands. Percentage shrub cover and snag basal area were similar in logged and mature stands (Table 2).

Bird Abundance and Species Richness

A total of 79 bird species was observed on the study sites, of which 76 were observed in logged stands and 37 in mature stands. The Veery (*Catharus fuscescens*), Black-throated Blue Warbler (*Dendroica caerulescens*), and Rose-breasted Grosbeak (*Pheucticus ludovicianus*) were the only species found exclusively in mature forest stands. Logged stands had statistically higher mean bird abundance and species richness than mature forest stands despite the fact that two 1-year old clearcut stands had the lowest density and species richness of all stands during any given year (Fig. 1). Logged stands had approximately twice the mean bird abundance (5.79 ± 0.79) and 1.5X the species richness (21.79 ± 1.20), on average, than that of mature forest stands (abundance 2.83 ± 0.27 ; richness 15.24 ± 1.16) (Fig. 1). The Red-eyed Vireo (*Vireo olivaceus*) was, by far, the most common bird species observed in this study and occurred with almost identical abundance in both logged and mature forest stands (Table 3). The Scarlet Tanager (*Piranga olivacea*) and Tufted Titmouse (*Baeolophus bicolor*) were also commonly observed and had a similar abundance in both logged and mature forest stands (Table 3). These two were the only bird species observed in all 18 study stands during any given year.

Table 2. Mean percentage vegetation cover (± 1 standard error and range) at different forest layers, woody debris index, and snag basal area on recently-logged and mature forest stands in Franklin, Patrick, and Henry counties, Virginia.

Variable	Logged	Unlogged
Overstory cover (%)	27.0 \pm 21.7 (0-55)	74.3 \pm 15.0 (54-96)
Midstory cover (%)	35.0 \pm 29.5 (0-84)	75.0 \pm 12.1 (55-92)
Shrub layer cover (%)	28.0 \pm 14.8 (7-53)	25.1 \pm 16.4 (5-57)
Herbaceous layer cover (%)	51.3 \pm 26.1 (13-78)	18.3 \pm 11.5 (5-46)
Woody debris index	199.4 \pm 51.1 (136-282)	68.2 \pm 27.6 (18-107)
Snag basal area (m ² /ha)	0.16 \pm 0.05 (0.05-0.45)	0.16 \pm 0.02 (0.10-0.28)

Overstory cover included vegetation cover >10 m tall; midstory cover included vegetation cover >2 m tall and <10 m tall; shrub layer cover included vegetation cover >0.5 m tall and <2 m tall; herbaceous layer cover included vegetation cover <0.5 m tall. Means are presented with ± 1 standard deviation. Minimum and maximum values are included in parentheses.

Relatively common species with higher abundances on mature forest stands compared to logged stands included the Ovenbird (*Seiurus aurocapilla*), Wood Thrush (*Hylochicla mustelina*), and Acadian Flycatcher (*Empidonax virens*). Other less abundant species showing a preference for mature stands included the Blue-headed Vireo (*Vireo solitarius*) and Yellow-billed Cuckoo (*Coccyzus erythrophthalmus*).

A long list of species more common on logged sites compared to mature forest stands included American Redstart (*Setophaga ruticilla*), Mourning Dove (*Zenaid macroura*), Carolina Wren (*Thryothorus ludivicianus*), Blue-gray Gnatcatcher (*Poliophtila caerulea*), Yellow-throated Vireo (*Vireo flavifrons*), American Goldfinch (*Carduelis tristis*), Indigo Bunting (*Passerina cyanea*), Eastern Towhee (*Pipilio erythrophthalmus*), Yellow-breasted Chat (*Icteria virens*), Red-bellied Woodpecker (*Melanerpes carolinus*), Brown Thrasher (*Toxostoma rufum*), Gray Catbird (*Dumetella carolinensis*), Hooded Warbler

(*Wilsonia citrina*), Kentucky Warbler (*Oporornis formosus*), Prairie Warbler (*Dendroica discolor*), and Chipping Sparrow (*Spizella passerina*) (Table 3).

DCA ordination axis 1 explained 51% of the variation in bird species composition among stands and axis 2 explained 23% of the variation. Recently logged stands were loaded on the right side of axis 1, with the most intensively logged stands appearing at the extreme end of the axis, and unlogged stands were clustered mostly to the left of axis 1 close to stands which had been partially cut (Fig. 2). Species on the ordination most closely placed with intensively logged stands included early-successional species such as Indigo Bunting, American Goldfinch, Prairie Warbler, and Gray Catbird. Bird species most closely associated with the left side of axis 1 included those typical of mature forests, such as Yellow-billed Cuckoo, Ovenbird, Acadian Flycatcher, Wood Thrush, and Blue-headed Vireo. Axis 2 is more difficult to interpret, although it may have been related to habitat patchiness, because two lightly selectively logged stands with a large amount of residual basal area interspersed with canopy gaps were isolated at the upper end of this axis (Fig. 2). Species most closely associated with these stands included Kentucky Warbler, American Redstart, Hooded Warbler, and White-breasted Nuthatch.

Relation of Bird Abundance and Species Richness to Habitat Variables

Bird species abundance was negatively correlated to overstory tree cover (Fig. 3) and midstory tree cover and positively correlated to herbaceous layer cover and large woody debris (Table 4). Bird species abundance was not significantly correlated with shrub cover. Relationships for species richness were similar except that species richness was not significantly correlated with overstory cover (Table 4). Similar results were obtained by using correlations of habitat variables with DCA ordination axes. DCA axis 1 was significantly and

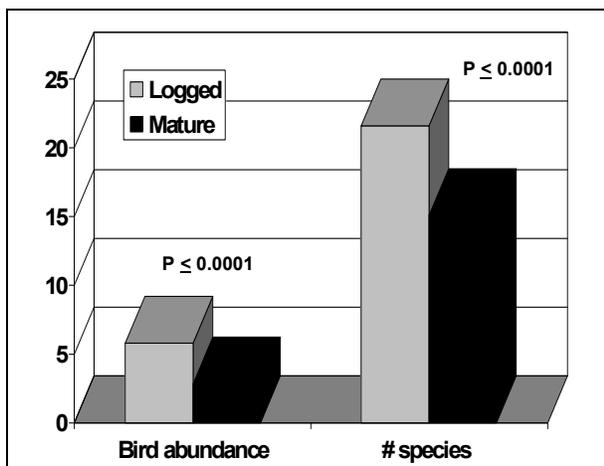


Fig. 1. Mean bird abundance (# birds/ha ± 1 standard error) and species richness (# bird species ± 1 standard error) on recently-logged and mature forest stands in Franklin, Patrick, and Henry counties, Virginia during 2004-2006.

Table 3. Mean bird density (# birds/ha ± 1 standard error) of bird species observed during point counts in recently-logged and mature forest stands in Franklin, Patrick, and Henry counties, Virginia during 2004-2006. An asterisk (*) indicates that species were observed on unlogged or logged stands at some point during the study although they were not detected during point counts.

Species	Mature forest	Recently-logged	Species	Mature forest	Recently-logged
Wild Turkey (<i>Meleagris gallopavo</i>)	0.02 (0.02)	0.02 (0.02)	White-eyed Vireo (<i>Vireo griseus</i>)	0.00 (0.00)	0.05 (0.03)
Accipiter Species (<i>Accipiter</i> sp.)	0.00 (0.00)	0.00 (0.00)*	Blue-headed Vireo (<i>Vireo solitarius</i>)	0.07 (0.05)	0.03 (0.02)
Northern Harrier (<i>Circus cyaneus</i>)	0.00 (0.00)	0.00 (0.00)*	Warbling Vireo (<i>Vireo gilvus</i>)	0.00 (0.00)	0.00 (0.00)*
Red-shouldered Hawk (<i>Buteo lineatus</i>)	0.00 (0.00)	0.00 (0.00)*	Black-and-white Warbler	0.02 (0.02)	0.03 (0.01)
Broad-winged Hawk (<i>Buteo playpterus</i>)	0.00 (0.00)	0.00 (0.00)*	(<i>Mniotilta varia</i>)		
Red-tailed Hawk (<i>Buteo jamaicensis</i>)	0.00 (0.00)	0.02 (0.01)	Black-throated Blue Warbler	0.01 (0.01)	0.00 (0.00)
Mourning Dove (<i>Zenaida macroura</i>)	0.00 (0.00)	0.06 (0.05)	(<i>Dendroica caerulescens</i>)		
Yellow-billed Cuckoo	0.03 (0.03)	0.00 (0.00)*	Yellow-rumped Warbler	0.06 (0.04)	0.10 (0.06)
(<i>Coccyzus americanus</i>)			(<i>Dendroica coronata</i>)		
Black-billed Cuckoo	0.00 (0.00)	0.00 (0.00)*	Chestnut-sided Warbler	0.00 (0.00)	0.00 (0.00)*
(<i>Coccyzus erythrophthalmus</i>)			(<i>Dendroica pensylvanica</i>)		
Ruby-throated Hummingbird	0.01 (0.01)	0.02 (0.01)	Pine Warbler (<i>Dendroica pinus</i>)	0.00 (0.00)	0.02 (0.02)
(<i>Archilochus colubris</i>)			Prairie Warbler (<i>Dendroica discolor</i>)	0.00 (0.00)	0.06 (0.04)
Red-headed Woodpecker	0.00 (0.00)	0.00 (0.00)*	Palm Warbler (<i>Dendroica palmarum</i>)	0.00 (0.00)	0.02 (0.02)
(<i>Melanerpes erythrocephalus</i>)			Yellow Warbler (<i>Dendroica petechia</i>)	0.00 (0.00)	0.00 (0.00)*
Red-bellied Woodpecker	0.03 (0.03)	0.12 (0.05)	American Redstart	0.01 (0.01)	0.40 (0.23)
(<i>Melanerpes carolinus</i>)			(<i>Setophaga ruticilla</i>)		
Pileated Woodpecker	0.01 (0.01)	0.00 (0.00)*	Blue-winged Warbler	0.00 (0.00)	0.00 (0.00)*
(<i>Dryocopus pileatus</i>)			(<i>Vermivora pinus</i>)		
Northern Flicker (<i>Colaptes auratus</i>)	0.01 (0.01)	0.03 (0.03)	Worm-eating Warbler	0.02 (0.01)	0.01 (0.01)
Downy Woodpecker	0.05 (0.03)	0.04 (0.01)	(<i>Helminthos vermivorus</i>)		
(<i>Picoides pubescens</i>)			Canada Warbler (<i>Wilsonia canadensis</i>)	0.00 (0.00)	0.00 (0.00)*
Hairy Woodpecker (<i>Picoides villosus</i>)	0.02 (0.01)	0.03 (0.02)	Hooded Warbler (<i>Wilsonia citrina</i>)	0.06 (0.02)	0.12 (0.04)
Eastern Kingbird (<i>Tyrannus tyrannus</i>)	0.00 (0.00)	0.00 (0.00)*	Kentucky Warbler	0.00 (0.00)*	0.09 (0.05)
Great-crested Flycatcher	0.02 (0.02)	0.02 (0.01)	(<i>Oporornis formosus</i>)		
(<i>Myiarchus crinitus</i>)			Common Yellowthroat	0.00 (0.00)	0.02 (0.02)
Eastern Phoebe (<i>Sayornis phoebe</i>)	0.01 (0.01)	0.05 (0.03)	(<i>Geothlypis trichas</i>)		
Eastern Wood-Pewee (<i>Contopus virens</i>)	0.00 (0.00)*	0.04 (0.02)	Yellow-breasted Chat (<i>Icteria virens</i>)	0.00 (0.00)	0.22 (0.13)
Least Flycatcher (<i>Empidonax minimus</i>)	0.00 (0.00)	0.00 (0.00)*	Northern Waterthrush	0.00 (0.00)	0.00 (0.00)*
Acadian Flycatcher	0.10 (0.05)	0.04 (0.04)	(<i>Seiurus noveboracensis</i>)		
(<i>Empidonax virescens</i>)			Louisiana Waterthrush	0.00 (0.00)	0.01 (0.01)
Willow Flycatcher (<i>Empidonax traillii</i>)	0.00 (0.00)	0.00 (0.00)*	(<i>Seiurus motacilla</i>)		
American Crow	0.05 (0.03)	0.00 (0.00)*	Ovenbird (<i>Seiurus aurocapilla</i>)	0.25 (0.07)	0.12 (0.09)
(<i>Corvus brachyrhynchos</i>)			Brown-headed Cowbird	0.00 (0.00)	0.03 (0.01)
Blue Jay (<i>Cyanocitta cristata</i>)	0.07 (0.03)	0.05 (0.02)	(<i>Molothrus ater</i>)		
Carolina Chickadee	0.07 (0.04)	0.11 (0.03)	Common Grackle (<i>Quiscalus quiscula</i>)	0.00 (0.00)	0.00 (0.00)
(<i>Poecile carolinensis</i>)			Orchard Oriole (<i>Icterus spurius</i>)	0.00 (0.00)	0.03 (0.01)
Tufted Titmouse (<i>Baeolophus bicolor</i>)	0.17 (0.05)	0.18 (0.05)	Baltimore Oriole (<i>Icterus glabula</i>)	0.00 (0.00)	0.00 (0.00)*
White-breasted Nuthatch	0.07 (0.03)	0.08 (0.05)	Summer Tanager (<i>Piranga rubra</i>)	0.00 (0.00)	0.00 (0.00)*
(<i>Sitta carolinensis</i>)			Scarlet Tanager (<i>Piranga olivacea</i>)	0.21 (0.09)	0.19 (0.10)
House Wren (<i>Troglodytes aedon</i>)	0.00 (0.00)	0.00 (0.00)*	Northern Cardinal	0.09 (0.03)	0.36 (0.14)
Carolina Wren	0.05 (0.02)	0.13 (0.03)	(<i>Cardinalis cardinalis</i>)		
(<i>Thryothorus ludovicianus</i>)			American Goldfinch (<i>Carduelis tristis</i>)	0.00 (0.00)	0.28 (0.09)
Ruby-crowned Kinglet	0.00 (0.00)	0.00 (0.00)*	Blue Grosbeak (<i>Guiraca caerulea</i>)	0.00 (0.00)	0.00 (0.00)*
(<i>Regulus calendula</i>)			Indigo Bunting (<i>Passerina cyanea</i>)	0.00 (0.00)*	0.38 (0.11)
Blue-gray Gnatcatcher	0.01 (0.01)	0.16 (0.05)	Rose-breasted Grosbeak	0.03 (0.03)	0.00 (0.00)
(<i>Poliptila caerulea</i>)			(<i>Pheucticus ludovicianus</i>)		
Brown Thrasher (<i>Toxostoma rufum</i>)	0.01 (0.01)	0.08 (0.06)	Eastern Towhee	0.00 (0.00)*	0.33 (0.10)
Gray Catbird (<i>Dumetella carolinensis</i>)	0.00 (0.00)	0.07 (0.04)	(<i>Pipilo erythrophthalmus</i>)		
Eastern Bluebird (<i>Sialis sialis</i>)	0.00 (0.00)	0.02 (0.02)	White-throated Sparrow	0.00 (0.00)	0.00 (0.00)*
American Robin (<i>Turdus migratorius</i>)	0.00 (0.00)	0.01 (0.01)	(<i>Zonotrichia albicollis</i>)		
Wood Thrush (<i>Hylocichla mustelina</i>)	0.26 (0.06)	0.12 (0.04)	Chipping Sparrow (<i>Spizella passerina</i>)	0.00 (0.00)	0.10 (0.07)
Veery (<i>Catharus fuscescens</i>)	0.00 (0.00)*	0.00 (0.00)	Field Sparrow (<i>Spizella pusilla</i>)	0.00 (0.00)	0.03 (0.02)
Cedar Waxwing	0.00 (0.00)	0.00 (0.00)*	Song Sparrow (<i>Melospiza melodia</i>)	0.00 (0.00)	0.02 (0.02)
(<i>Bombcilla cedrorum</i>)			Unidentified Birds	0.12 (0.04)	0.31 (0.10)
Red-eyed Vireo (<i>Vireo olivaceus</i>)	0.86 (0.11)	0.85 (0.18)			
Yellow-throated Vireo	0.01 (0.01)	0.10 (0.06)			
(<i>Vireo flavifrons</i>)					

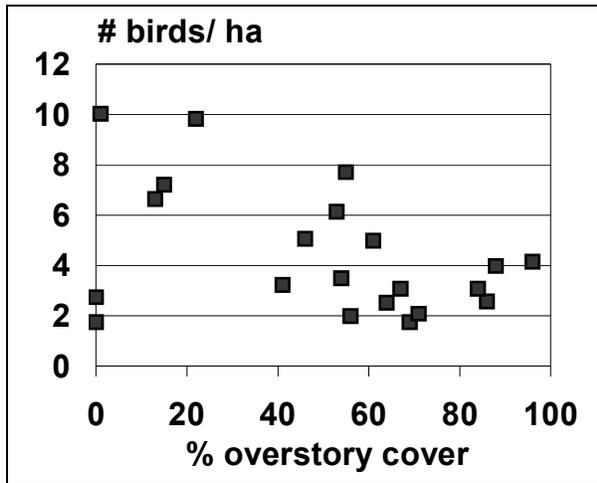


Fig. 3. Relationship between bird abundance and percentage overstory cover on recently-logged and mature forest stands in Franklin, Patrick, and Henry counties, Virginia during 2004-2006. The two stands at the bottom left of the chart are recent (<1 year-old) clearcut stands.

1990). While common in intensively logged stands as well, other species occurred commonly in selectively logged forest, including the Carolina Wren, Northern Cardinal, Blue-gray Gnatcatcher, and Yellow-throated Vireo. Interestingly, the Red-bellied Woodpecker (*Melanerpes carolinus*) was four times more abundant in logged stands compared to mature forest stands, although the abundances of both Downy (*Picoides pubescens*) and Hairy Woodpeckers (*P. villosus*) were similar among logged and mature forest stands.

As expected, Ovenbird and Wood Thrush were much more abundant on mature forest stands than logged stands and rarely occurred in intensively cut stands. Both of these species have been described as forest interior species (Ehrlich et al., 1988). Blue-headed Vireo and Acadian Flycatcher were also more common in mature stands than logged stands, but they did occur in some selectively logged stands. Interestingly, other birds considered to be forest interior species, including Scarlet Tanager and Red-eyed Vireo, had similar abundances on logged and mature forest stands. These species were even retained on intensively logged stands provided that there were small patches of mature trees left within the stand, such as along riparian areas.

Bird species abundance was negatively correlated with overstory tree cover and midstory tree cover and positively correlated with herbaceous layer cover and coarse woody debris. It is difficult to determine the relative importance of these relationships for habitat quality, however, because they are autocorrelated to

some degree. Logging decreases overstory cover, thereby increasing light availability to the forest floor, which stimulates the growth of understory cover. The upper portions of trees and non-merchantable trees that are felled or toppled during the logging operation increase coarse woody debris on the forest floor, unless they are chipped, burned or otherwise disposed of following logging. Removal of woody debris occurred on only one stand in this study. Herbaceous layer cover and woody debris provide foraging habitat and cover for many species of birds (Titterton et al., 1979; Hagan et al., 1995).

It is not intuitively clear why midstory tree cover would be negatively correlated with bird abundance or species richness. Many intensively logged stands result in low overstory tree cover, but relatively high midstory cover. The retention of these trees may also provide perching, singing, cover, and foraging habitat for birds. On the other hand, an abundance of midstory cover may suppress herbaceous layer cover, which appears to be important for many bird species. It is also interesting that shrub layer cover was not correlated with bird abundance or species richness in this study given that many shrub and tree species respond positively to increased light availability following logging (Carter & Fredericksen 2007). Many unlogged stands in this study, however, had areas with relatively dense shrub cover consisting of Mountain Laurel (*Kalmia latifolia*) and Rhododendron (*Rhododendron* spp.).

There was little correlation between snag basal area and bird abundance or species richness in this study despite the fact that many studies find snag tree retention to be critical for retaining bird species richness in logged stands (Sallabanks & Arnett, 2005). I have personally observed, however, that loggers in the study area do not purposely fell snags, or other non-merchantable stems, although they may be toppled inadvertently by falling trees or logging machinery. Any loss of snag basal area due to toppling during the logging operation may perhaps be offset by the fact that some snags are likely to be created during logging due to root damage or other injuries to live trees. Indeed, mean snag basal area was identical in logged and mature forest stands (Table 2).

It is important to note several limitations of this study. First, the study only addressed bird species abundance during the early breeding season. Studies have documented habitat shifts during the breeding season for birds, particularly the use by fledglings of areas with dense vegetation, such as those created by logging, for bird species that tend to nest in mature forest (Anders et al., 1998; Vega Rivera et al., 1998; Marshall et al., 2003; King et al., 2006; Vitz & Rodewald, 2006). The impact on wintering habitat is

also unknown. The timing of this study also coincided with spring migration during which birds are less selective of habitat or perhaps use different habitats than they do during the breeding season. Finally, this study did not document bird productivity differences due to logging or the intensity of logging. The presence of a bird species on a site does not indicate that it is also reproductively successful on that site (Brawn & Robinson, 1996), although studies have documented increases in bird productivity following logging in eastern forests (Ross et al., 2001; Weakland et al., 2002). Despite these limitations, this study is one of only a very few studies documenting the relative abundance, species richness, and composition of bird communities on recently-logged or mature non-industrial private forestlands.

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